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生态学报

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封面图说:美丽的油松松枝——油松又称红皮松、短叶松。树高可达30m,胸径达1m。其树皮下部灰褐色,裂成不规则鳞块;针叶2针一束,暗绿色,较粗硬;球果卵形或卵圆形,长4—7cm,有短柄,与枝几乎成直角。油松适应性强,根系发达,树姿雄伟,枝叶繁茂,有良好的保持水土和美化环境的功能,是中国北方广大地区最主要的造林树种之一,在华北地区无论是山区或平原到处可见,人工林很多,一般情况下在山区生长最好。在山区生长的油松,多在阴坡、半阴坡,土壤湿润和较肥沃的地方。

彩图及图说提供:陈建伟教授 北京林业大学 E-mail: cites.chenjw@163.com

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氮沉降对森林土壤有机质和凋落物分解的影响及其微生物学机制

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摘要: 氮沉降持续增加背景下土壤 C:N:P 化学计量比和 pH 环境等的改变及其可能的土壤微生物学机制已经成为陆地生态系统与全球变化研究的新生长点和科学研究前沿。以生态化学计量学和土壤微生物生态学为理论基础, 综述了氮沉降对森林土壤有机质和凋落物分解的影响及其微生物学机制的基本理论、最新进展、研究热点与难点, 旨在促进全球变化背景下陆地生态系统地下生态学的研究。氮沉降持续增加会导致森林生态系统磷循环加速, 导致磷限制。氮沉降不但改变森林土壤有机质和凋落物的 C:N:P 化学计量比和降低土壤 pH 值, 而且改变土壤微生物生物量碳氮磷、细菌、真菌和放线菌的组成以及影响碳氮磷分解的关键酶活性。氮沉降对森林土壤有机质和凋落物分解的影响表现为促进、抑制和无影响, 其影响的差异可能来源于微生物效应的不同。叶片在凋落前有显著的氮磷养分回收, 但是根无明显的养分回收, 造成土壤有机质和凋落物的 C:N:P 化学计量比存在明显差异。基于 DNA/RNA 等分子生物学方法为土壤微生物生态学研究提供了强有力的手段, 将促进氮沉降对森林土壤有机质和凋落物化学计量比改变的微生物学机制研究。

关键词: 氮沉降; 磷限制; 分解; C:N:P 计量比; 土壤微生物

The effect of nitrogen deposition on forest soil organic matter and litter decomposition and the microbial mechanism

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Abstract: The changes of soil C:N:P ratios and pH under nitrogen deposition, its effect on soil organic matter and litter decomposition and the possible soil microbial mechanism are the new position and scientific research front of terrestrial ecosystem and global change. The paper based on ecological stoichiometry and soil microbial ecology, summarized the basic theory, new advances, hot and difficult points of research on the effect of nitrogen deposition on soil organic matter and litter decomposition and the microbial mechanism, to advance the research of terrestrial underground ecology under global change. The continuing increasing nitrogen deposition accelerated the phosphorus cycling of forest ecosystem, caused phosphorus limitation. Nitrogen deposition not only changed soil organic matter and litter C:N:P ratios and decreased soil pH, but also changed soil microbial components of carbon, nitrogen and phosphorus, composition of bacterium, fungus and actinomycetes, the key enzymes for decomposition of carbon, nitrogen and phosphorus. The effects of nitrogen deposition on soil organic matter and litter decomposition have three states: motivation, suppression, no effect, the reasons for the

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different effect maybe come from different microbial effect. Leaves had obvious resorption of nitrogen and phosphorus, but root had less resorption rate, which caused clearly different C:N:P ratios between soil organic matter. The molecular methods based on DNA/RNA provided advanced tools for soil microbial ecology research, which would promote researches on microbial mechanism of soil organic matter and litter decomposition.

Key Words: nitrogen deposition; phosphorus limitation, decomposition; C:N:P ratios; soil microbial

全球氮沉降持续增加(可能从现在的 $1.5 \text{ kgN} \cdot \text{hm}^{-2} \cdot \text{a}^{-1}$ 增加到 $4.2 \text{ kgN} \cdot \text{hm}^{-2} \cdot \text{a}^{-1}$)^[1],其对森林土壤碳汇的影响依然充满争议,其影响机制一直是陆地生态系统和全球变化研究的科学前沿。例如:氮沉降加速了南美热带雨林土壤碳释放^[2],但是增加了欧洲森林^[3]和北半球温带森林^[4]土壤碳吸收,也可能是促进我国南方成熟森林土壤碳吸收的原因^[5]。土壤有机质和凋落物的分解过程是决定森林土壤碳汇的核心关键过程^[7]。目前,人类活动产生并排放到大气中的活性氮已由 1860 年的 15 TgN/a 增加到 2000 年的 165 TgN/a ,超过了陆地自然生态系统的最大估计固氮量 130 TgN/a ^[16],其中 70% 通过干湿沉降到达地表^[17]。大气氮素沉降主要包括湿沉降和干沉降^[18]。硝态氮与铵态氮是氮沉降的两种主要形态,其中铵态氮占总沉降通量的 60% 以上。据预测,全球氮沉降速率到 2050 年将加倍,并有相当一部分地区超过 $50 \text{ kgN} \cdot \text{hm}^{-2} \cdot \text{a}^{-1}$ ^[19]。目前我国已经成为欧洲、北美之后的第三大氮沉降区 ($12.9 \text{ kgN} \cdot \text{hm}^{-2} \cdot \text{a}^{-1}$),氮沉降变化范围为 $1.0\text{—}74.3 \text{ kgN} \cdot \text{hm}^{-2} \cdot \text{a}^{-1}$,其中东南区域的氮沉降水平最高,高达 $35.6 \text{ kgN} \cdot \text{hm}^{-2} \cdot \text{a}^{-1}$,约每年增加 $0.34 \text{ kgN} \cdot \text{hm}^{-2} \cdot \text{a}^{-1}$,由东南向西北呈递减趋势,全国形成东南、四川盆地、长江中下游平原等高氮沉降中心^[20]。

氮沉降不但改变土壤 C:N:P 化学计量比^[10-11],而且会降低土壤 pH^[12],并加速土壤磷循环^[8]。全球已发表的氮添加实验的综合分析表明,几乎所有生态系统中都存在由于氮素添加而导致的磷限制^[9],进而影响土壤有机质和凋落物的分解过程。氮沉降持续增加背景下土壤微生物的生物量、组成与酶活性的改变是调控土壤有机质和凋落物分解的核心机制^[13-14]。氮、磷添加控制试验是调控土壤 C:N:P 生态化学计量比的有效手段^[15]。2011 年 *New Phytologist* 杂志组织“全球变化下陆地生态系统化学计量比的适应性 (Stoichiometric flexibility in terrestrial ecosystems under global change)”专题学术研讨会,指出氮沉降持续增加背景下土壤 C:N:P 化学计量比和 pH 环境等的改变对土壤有机质和凋落物分解的影响及其可能的土壤微生物学机制尚未清楚,已经成为陆地生态系统与全球变化研究的新生长点和科学研究前沿。

施氮实验证明氮添加能促使生物体产生更多的胞外磷酸酶,导致土壤有机质释放更多的磷酸盐^[21];在高氮沉降的区域也产生类似的磷循环情景^[22]。然而,氮沉降的持续增加已经促使生态系统由氮限制转变为磷限制^[23]。但是磷限制对生态系统碳固定的影响可能被远远低估^[24]。通常认为热带森林由于受磷限制而导致其氮沉降效应没有其它生态系统效果明显^[23]。全球海洋与非农业陆地生态系统磷限制的情况正在日益加强^[11]。随着大气 CO_2 浓度和氮沉降增加,全球森林生长与碳库并没有显著增加或降低^[25],这可能由于磷限制掩盖了 CO_2 浓度和氮沉降增加的潜在效应。随着氮沉降的持续增加,磷限制性会逐渐增强^[11]。人类活动导致其它元素(特别是氮)的大量增加是陆地生态系统存在磷限制的重要原因^[9, 23]。

1 氮沉降对土壤有机质和凋落物化学计量比的影响

1.1 土壤 C:N:P 化学计量比的生态学意义

生态化学计量学主要研究生态过程中化学元素的比例关系,目前主要集中于碳、氮和磷元素的计量关系^[26-27]。生态系统碳-氮-磷平衡的标志之一就是土壤微生物和植物需求之间利用土壤为平台,通过动态交换达到并维持相对平衡的碳、氮和磷元素生态化学计量比^[28]。生态系统中碳与关键养分元素(氮、磷)的生态化学计量比的差异能够调控和影响生态系统中碳循环过程^[29]。土壤 C:N:P 化学计量比综合了生态系统功能的变异性,是反映土壤碳-氮-磷循环的主要指标,有助于理解生态过程对全球变化的响应,成为确定土壤碳-氮-磷平衡特征的重要参数^[30]。植物叶片、根系、土壤有机质和凋落物中 C:N:P 化学计量比可以表征氮磷养

分限制状况^[31-32]。

土壤碳库的变化受到土壤微生物为维持自身碳/养分平衡需求的制约,这表明土壤碳排放过程是与外界氮磷输入相联系的^[33]。土壤 C:N 比直接控制硝化速率并对用于反硝化作用的硝酸根有间接影响,与土壤微生物对其的分解能力密切相关^[34-35]。土壤有机质和凋落物 C:N 比大于 25 时对微生物来说是氮限制性的,因此,氮首先被微生物固定利用而不是被硝化或反硝化。在全球尺度上,土壤 C:N 比能解释 99.2% 的森林溪流 DOC 年通量变化^[36]。土壤 N:P 比是影响微生物群落多样性的重要因子,也是一个土壤养分限制的指示因子。在区域水平上,氮或磷限制条件下,土壤 N:P 比会影响植物生产力与组成,同时会影响土壤微生物的组成与活性^[37-38],表明土壤和土壤微生物 C:N:P 化学计量比与土壤有机质和凋落物分解密切相关^[39-40]。土壤微生物 C:N、C:P 比可以作为相对于氮和磷的碳利用效率指标,而 C:P 比也可作为衡量土壤有机物质矿化释放磷的一种指标^[41]。

1.2 氮沉降对土壤 C:N:P 化学计量比和 pH 的影响

在全球尺度上,土壤中碳-氮-磷含量具有显著的正相关关系,但并不是线性关系,因为磷的增加量通常小于碳氮^[42]。全球及中国土壤的 C:N:P 化学计量比分别为 186:13:1 和 60:5:1^[42-43]。无论土壤中碳氮含量如何变化巨大,低磷含量总是导致高的 C:P 和 N:P 化学计量比,因此土壤 C:N:P 化学计量比的大小主要是受磷含量控制^[43]。

由于人类活动对自然生态系统输入物质的 C:N:P 化学计量比的改变,必将对自然生态系统植被、土壤与土壤微生物的 C:N:P 化学计量比产生影响^[44-45]。研究表明,自然过程中全球 C:N:P 化学计量比约为 20333:43:1,人类活动产生的 C:N:P 化学计量比约为 667:12:1^[46]。例如,氮添加会导致植物叶片中氮含量增加,进而降低了叶片 C:N 比^[47-48]。胞外酶通常能够显示土壤微生物的养分需求,在养分不受限制的情况下其 C:N:P 生态化学计量比是 1:1:1^[30]。由于养分可利用性与土壤类型相关,在磷限制的土壤中其通常显示更高的 C:P 比^[30]。随着土壤深度的变化,土壤磷的降低速率要低于土壤碳氮^[43],因此土壤 C:P 比随土壤深度的下降速度要高于 C:N 比。随着氮沉降的增加,一些研究已经观测到土壤 N:P 比的增加^[10]。全球而言,人类活动对生物圈投入的 N:P 比为 22.8—44.6:1,比大多数陆地植物最优生长条件下土壤 N:P 比大了 5%—100%^[46]。

氮沉降增加了土壤有机质和凋落物分解过程中的氮含量,改变了土壤有机质和凋落物的 C:N:P 化学计量比并降低土壤 pH^[49-50]。随着氮沉降的持续增加,土壤 pH 值也随之下落,这也降低了土壤的阳离子交换量,改变了土壤中盐基阳离子数量与种类^[51]。Matson 等证明大多数进入热带生态系统的氮流失到水体或大气,导致硝化速率增加,氮损失导致盐基阳离子损失并降低土壤 pH,这会反过来降低热带森林的碳固定能力^[52]。

1.3 氮磷添加对土壤有机质和凋落物分解的影响

氮沉降最直接的效应是改变土壤微生物分解底物的质量(C:N:P 化学计量比)、数量以及 pH 环境等,进而影响土壤微生物生物量、组成和酶活性,从而促进或限制土壤有机质和凋落物等分解速率。温带和北方林明显受到可利用氮的限制^[19, 53-54],而热带和亚热带森林处于有效氮富集状态,而存在磷限制^[19, 55]。Neff 等研究表明,氮添加显著加快了较轻的土壤碳分解,而使得较重的土壤碳化合物更为稳定^[56]。同时,氮沉降对土壤有机物质和凋落物分解的影响因分解阶段不同而表现出差异: NH_4^+ 和 NO_3^- 进入新鲜的、刚脱落的凋落物中会促进纤维素和可溶物质的初期分解,而相同化合物进入腐殖质(分解结束阶段)会显著抑制其活性^[57]。莫江明等发现在分解初期添加氮促进马尾松林凋落物的分解,而对荷木凋落物分解无显著影响,对季风常绿阔叶林凋落物分解的影响随树种不同而异,总体上氮添加对凋落物分解是抑制作用大于促进作用,并认为少量土壤可用性氮增加可以提高凋落物分解速率,但过高则抑制凋落物分解^[58]。Cao 等在南亚热带的马占相思和桉树人工林的氮磷添加试验则表明氮磷添加均促进了马占相思林的土壤呼吸,但是降低了桉树林的土壤呼吸^[59]。Zhang 等^[60]在会同地区的杉木凋落物与土壤加氮试验表明,杉木土壤碳稳定依赖于凋落物和氮的

输入。

土壤有机质和凋落物的氮磷含量会影响土壤微生物对土壤有机质和凋落物的分解:当氮含量低时,土壤微生物会降低其单位氮的碳利用效率。叶片在凋落前有显著的氮磷养分回收,但根没有或较少,因此死根中的氮磷养分含量远高于叶凋落物。土壤有机质和凋落物的 C:N:P 化学计量比存在明显差异,而土壤有机质库具有较高的氮磷养分含量,供应氮磷养分的速率高^[61-62]。土壤有机质和凋落物的碳氮磷化学计量比是其分解的重要决定因子,然而元素的化学计量比对其分解过程中微生物的氮磷周转影响现在研究不足。例如山毛榉凋落物放在培养 3 个月及 6 个月后分解者氮磷循环发生改变。总蛋白分解、氮矿化(氨化)、硝化速率与凋落物的 C:N 比存在负相关。磷的矿化速率与凋落物的 C:P 比存在负相关,凋落物 C:N 比与无机氮循环的负相关关系比其与总蛋白分解的负相关关系更强,显示底物的化学计量比对细胞内过程的效应要比胞外酶催化过程的效应要强。胞外蛋白分解主要受底物的可利用性限制而较少受到蛋白酶数量的限制。氮库与磷库是密切正相关而且其预期产量与消费过程指出其控制微生物、凋落物的氮磷循环。凋落物 C:N 比与磷酸酶活性的负相关关系显示微生物倾向于分配足够的碳和养分来生产胞外酶来增加养分的供应。资源的碳氮磷化学计量比对微生物的分解凋落物过程的氮磷循环具有很强的效应,氮磷的矿化与保持凋落物微生物群落的细胞内稳性具有较强的耦合作用^[63]。

氮输入与变暖会通过改变微生物活性和分解过程而改变土壤碳储量。施肥可以增加大多数土壤碳含量,降低异养呼吸而减少碳损失。但是土壤碳库的增加在不同碳库并不一致,活性的土壤碳库在施肥的情况下降解非常快,但是缓效的碳库却延长了其周转时间。土壤碳库对氮输入的变化与两类微生物胞外酶相关。小的活跃碳库会增加水解酶活性,而周转时间长的缓慢碳库会降低氧化酶活性,水解酶在降解施肥土壤中更复杂的碳组分。增温会整体上增加土壤呼吸,氮输入会显著增加缓效碳库的温度敏感性。氮沉降有增加热带森林土壤碳库的趋势,但是增加的碳库是否能长期稳定可能取决于将来的温度变化^[64]。

2 氮沉降对土壤微生物组成与酶活性的影响

2.1 土壤微生物组成与酶活性的测定方法

土壤微生物组成与酶活性是反应土壤有机质和凋落物分解过程的重要指标^[6, 64, 66]。土壤中生活着大量的微生物(每克土壤可含 10^9 细菌、 10^7 放线菌和 10^6 真菌),它们是土壤生物地球化学循环的驱动者^[29, 67-68]。目前,土壤微生物生物量碳氮磷的测定方法是熏蒸培养法^[69-70]。土壤微生物细菌、真菌和放线菌组成与活性主要依靠磷脂脂肪酸(phospholipid fatty acid, PLFA)分析技术^[71-74]。目前国际上研究较多的土壤酶活性是土壤胞外酶活性,多采用微孔板法分析土壤的酶活性^[75],而我国酶活性分析多采用传统的培养、比色法^[70, 76]。值得注意的是,基于 DNA/RNA 等分子生物学方法为土壤微生物生态学研究提供了强有力的手段,如 RT-PCR、PCR-DGGE,第二代基因组测序技术^[77]。

2.2 氮磷添加对微生物组成和酶活性的影响

在研究氮、磷添加对土壤微生物组成和酶活性影响时,植被类型、土壤有机质含量等特征不容忽视。不同土地利用类型土壤微生物与酶活性的差别与土壤有机质含量密切相关,在高有机质含量的土壤中土壤酶活性较高,反之亦然^[78];在不同土地利用类型中,土壤酶活性、真菌:细菌的比和土壤质地与土壤有机质化学组分密切相关^[68]。在内蒙古呼伦贝尔草原研究也表明,资源可利用性(植物生物量、土壤含水量、土壤 N:P 化学计量比)是决定土壤微生物功能多样性的主要因子^[38]。Zechmeister-Boltenstern 研究了欧洲 12 种森林植被类型和氮沉降水平下土壤微生物状况,结果表明植被类型对土壤微生物种群结构影响明显,仅当氮沉降水平较高($>30 \text{ kgN} \cdot \text{hm}^{-2} \cdot \text{a}^{-1}$)时,氮沉降的影响才掩盖了植被类型的效应^[14]。

氮添加是影响微生物细菌、真菌和放线菌组成的重要因素^[79-80]。在北美碳和氮含量低的土壤中,氮添加主要增加放线菌的生物量^[81]。在北方硬阔林的土壤中,氮添加降低菌根真菌的生物量^[13]。美国北部阔叶林 12a 氮沉降导致土壤微生物生物量和丛枝菌根真菌生物量降低 24%—36%,并导致土壤真菌与细菌的比例降低了 10%^[13]。Treseder 综合评价了 82 个田间施氮试验对土壤微生物的影响,结果表明微生物生物量平均降

低约 15%,且土壤 CO₂ 排放减少;长期的高施氮试验中细菌和真菌有降低的趋势^[80]。以细菌为主的土壤中,氮素增加抑制了难分解有机质的降解^[75]。瑞典北部北方森林环割和 34a 施氮试验研究表明,环割和施氮均导致土壤真菌生物量降低 45%,说明氮沉降可能同环割处理相同,降低了对外生菌根真菌供碳量,而细菌生物量随自然梯度的 pH 增加而明显增加^[79]。在低海拔的热带森林中,氮沉降导致革兰氏阴性细菌生物量增加,活性有机碳化学组分流失增加;而在高海拔的热带森林中,氮沉降导致真菌增加,降低了有利于土壤有机碳储存的化学组分^[82]。对美国 28 种土壤进行 1a 室内增氮培养试验研究表明,增氮改变了土壤微生物群落组成,增加了放线菌和硬壁菌生物量,降低了酸杆菌和疣微菌生物量^[81]。

氮沉降增加会加速土壤磷循环并导致磷限制,改变土壤 C:N:P 化学计量比和 pH 环境,这将导致土壤微生物组成与酶活性发生变化^[8]。Cleveland^[42]研究表明,不同于土壤 C:N:P 化学计量比在全球的一致性,土壤微生物 C:N:P 化学计量比随着植被类型变化有明显区别。例如森林土壤微生物 C:P、N:P 比明显高于草原土壤微生物 C:P、N:P 比,主要由于森林土壤微生物的磷含量低^[42]。土壤 N:P 比是影响微生物群落多样性的重要因子,也是一个土壤养分限制的指示因子。在氮或磷限制环境下,土壤 N:P 比能够显示其对植物生产力与组成的改变来影响土壤微生物群落多样性。通过氮沉降、施肥、增温等产生的土壤 C:N:P 化学计量比的改变均可能影响土壤微生物的分布^[83-84]。真菌的碳利用效率与 C:N、C:P 比有正相关关系,而细菌的碳利用效率和 C:N、C:P 比是负相关关系,而且这种关系在不同生物量条件下保持稳定。这显示真菌比细菌有更高的碳需求^[80]。

磷的添加改变微生物生物量与组成,可以增加土壤有机质和凋落物分解^[38, 85]。例如鼎湖山森林 3a 施磷试验表明,亚热带老龄森林中施磷促进了土壤微生物生物量的增加,增加了土壤微生物中细菌和真菌的量,并增加了土壤的真菌与细菌的比^[86]。在磷限制区域,细菌生长非常缓慢,并由于 RNA 浓度的降低导致其磷含量也降低^[87-88]。

不同氮沉降处理对土壤酶的活性影响差异较大。氮沉降降低了凋落物表面酶活性效能,并且使酶的功能从催化氮素转化向催化磷素转化转变,并且从多酚氧化酶向多聚糖水解酶转变^[67]。长期施氮对温带阔叶林地表凋落物酶活性的影响超过了对土壤酶活性的影响,施氮增加了凋落物和土壤中脲酶、酸性磷酸酶、糖苷酶活性;增加了凋落物中酚氧化酶的活性,但是降低了土壤中酚氧化酶的活性^[75]。氮添加显著抑制矿质土壤中 β-1,4-葡糖苷酶的活性,而且抑制了地表凋落物的酚氧化酶活性^[89]。美国的 28 种土壤室内 1a 增氮培养试验研究表明,增加氮显著降低了葡萄糖苷酶、酸性磷酸酶、亮氨酸胺酶、过氧化物酶等土壤微生物的酶活性^[81]。对于氮限制的 3 个温带森林,经过 1a 的施氮后土壤苯酚氧化酶、过氧化物酶有增加,也有降低的趋势,不同生态系统对酶活性具有不同响应特征^[90]。Marklein 和 Houlton 利用全球 34 个自然生态系统氮磷添加试验数据综合分析了氮、磷添加对土壤磷酸酶活性的影响,氮添加增加了土壤磷酸酶的活性,而磷添加抑制了土壤磷酸酶活性^[8]。植物和土壤把过剩的氮素分配给磷酸酶,延迟氮沉降增加而导致磷限制。Sinsabaugh 等综合全球 40 个生态系统的土壤酶活性数据研究表明,β-1,4-葡糖苷酶,β-1,4-N-乙酰葡糖胺酶和磷酸酶活性会随着土壤有机质浓度的增加而增加,而亮氨酸胺酶的活性与土壤有机质浓度无明显关系^[30]。不同酶在氮沉降影响下随时间的动态变化无统一规律,蔗糖酶波动较大,而脲酶和过氧化氢酶两者趋势较一致等。

土壤 pH 环境改变直接影响土壤微生物组成和酶活性变化^[91]。在北美洲和南美洲洲际尺度区域调查研究表明土壤细菌种群多样性和丰富度主要受土壤 pH 值影响,在酸性土壤中微生物多样性最低,在中性土壤中微生物多样性最高^[92]。在全球尺度利用 40 个生态系统调查结果,发现 β-1,4-葡萄糖苷酶、纤维素水解酶、β-1,4-N-乙酰葡糖胺酶、磷酸酶活性会随着土壤有机质含量的增加而增加,而亮氨酸胺酶、苯酚氧化酶、过氧化物酶的活性与土壤有机质浓度无明显关系,但是这 7 种酶活性都与土壤 pH 显著相关,而与年均温或年均降雨量相关性不明显^[30]。

3 总结和展望

我国为世界上第三大氮沉降区,氮沉降持续增加而导致了森林生态系统磷循环加速而导致磷限制。氮沉

降改变森林土壤有机质和凋落物的 C:N:P 化学计量比和降低土壤 pH 值,而且改变土壤微生物生物量碳氮磷、细菌、真菌和放线菌的组成以及影响碳氮磷分解的关键酶活性。氮沉降持续增加背景下土壤 C:N:P 化学计量比和 pH 环境等的改变对土壤有机质和凋落物分解的影响及其可能的土壤微生物学机制尚未清楚。以往研究多独自开展生态化学计量学或土壤微生物生态学研究,迫切需要开展生态化学计量学和土壤微生物生态学交叉的综合研究,促进全球变化背景下我国陆地生态系统地下生态学的研究。

氮沉降对土壤有机质和凋落物化学计量比改变的作用机理存在着多种解释,但目前仍无普遍接受的结论^[93-94]。现有研究多数以针叶林为研究对象在氮沉降严重的欧洲和北美森林中进行,结果带有局限性。热带和亚热带地区的森林类型比较复杂,氮循环在热带地区与温带地区之间存在差别,磷是热带亚热带地区重要的限制因子,但是在温带的限制效应并不明显。如何利用生态化学计量特征来合理比较不同区域森林土壤有机质和凋落物的分解特征及其对氮沉降的响应?土壤有机质和凋落物的化学计量比对细胞内过程的效应和胞外酶催化的过程的效应是否与林型有关^[63],其氮磷的矿化与保持微生物群落的细胞内稳性的耦合如何协调^[64]?氮输入与土壤响应参数之间的关系在森林表层与矿质土壤氮矿化在空间的差异大,而且森林表层土壤碳氮比和矿质土壤碳储量存在十来倍的差异。空间格局上的差异可能来源于是否仅仅是土壤有机质分解与稳定的差异,以及活性碳库与惰性碳库的不同响应均值得深入研究^[65]。

资源的化学计量比对消费者的活性和数量均是重要的控制因子。凋落物的化学成分是分解者的主要食物来源,也是分解者活性的主要驱动因子。理论预测资源生态化学计量比对消费者的数量有高度控制,保持严格的内稳性,由于需要消耗养分来保持消费者身体组织的平衡。分解者对养分的获取通常与资源化学计量比的不平衡有关。通过对纯林凋落物的研究表明:土壤与凋落物的化学计量比之间没有联系,微生物活性与土壤化学计量比有关^[95],但是作为分解者的微生物能否在分解过程中保持内稳性尚没有结论。基于 DNA/RNA 等分子生物学方法为土壤微生物生态学研究提供了强有力的手段,将促进氮沉降对森林土壤有机质和凋落物分解的微生物学机制研究。如 RT-PCR 技术可以定量研究氮沉降对土壤氮循环的关键基因(固氮基因 *nifH*、氨氧化基因 *amoA*、氮还原基因 *nirS*、*nirK*、氧化亚氮还原基因 *nosZ*)的影响^[96]。PCR-DGGE 技术作为一种高通量的测度微生物群落结构和多样性的方法,可以用来探讨不同的环境因子与微生物群落结构和多样性的关系。常以基于 rDNA 指纹图谱的可操作分类单元(OTU)的数目表征物种丰富度^[97]。Lekberg^[98]在《Nature》撰文指出 454 公司推出的基于焦磷酸测序法的超高通量基因组测序系统(Genome Sequencer 20 System),为微生物的多样性和丰度等群落结构分析提供了丰富的数据支持。结合¹³C 核磁共振技术,可以对不同组分有机碳的变化机理进行深入研究,结合根窗等原位观测技术,可以对原位土壤酶活性分布的热点区域进行根分布特征间关系进行深入研究^[82]。

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