

城市污泥应用于陆地生态系统研究进展

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摘要:随着城市化进程加快和人口剧增,城市污泥已成为世界许多城市面临的主要环境问题之一,且不合理的管理可引发严重的环境污染。城市污泥含有大量、微量元素和有机质,可能对土壤及其生产力有利,特别对退化土壤能进行有机修复,并改善土壤理化特性,譬如土壤结构和营养含量。目前,污泥作为一种有机肥料已成为普遍措施。但来自工业及生活污水的污泥常含有重金属、病原物及有毒有机物等,潜在的毒性可能对生态系统构成危险,因此必须经过污泥预处理才可安全施用。评述了近年来国内外污泥应用于陆地生态系统 4 个方面的主要研究进展:(1)污泥处理与处置方法;(2)污泥应用于农田、草地及森林生态系统;(3)污泥对土壤生态系统的影响,包括污泥对土壤理化性质、土壤酶及微生物的影响;(4)污泥应用的环境效应。提出未来我国污泥处置及利用需要重视的研究领域和方向。

关键词:污泥; 陆地生态系统; 土壤生态系统; 环境效应

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Research progress of the application of urban sewage sludge to terrestrial ecosystems

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Abstract: Due to rapid population growth and urbanization, urban sewage sludge has become one of the most important environmental problems in many cities in the world, and improper management would cause serious environmental issues. Generally, sewage sludge contains macro and micro elements and organic matter, which are beneficial to soils and their productivity. In particular, sewage sludge can provide useful organic amendment in degraded soils to improve soil chemical and physical properties, such as soil structure and nutrient content. Nowadays, the use of sewage sludge as an organic fertilizer has become a common practice, however, sewage sludge derived from both industrial and domestic sources can often contain considerable amounts of heavy metals, pathogens and toxic organics, and the persistence and potential toxicity of these substances may pose a risk to ecosystems. Hence, the utilization of the sewage sludge safely has to be ensured by the preliminary treatment of sewage sludge. The recent advances of the application of sewage sludge to terrestrial ecosystems within four processes were reviewed as follow: (1) the sludge treatment and disposal technologies; (2) the application of sewage sludge to agricultural ecosystems, grassland ecosystems and forest ecosystems; (3) the effects of sewage sludge on soil ecosystem, including soil chemical and physical properties, soil enzymes and soil microorganisms; (4) the environmental effects of sewage sludge application. Finally, some important research fields and trends related to the disposal and application of sewage sludge in the future were also proposed.

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城市污泥是指城市生活污水、工业废水处理过程中产生的固体废弃物。因人口快速增长和城市化,不合理的污泥废弃物贮存对城市环境构成威胁,并日益成为世界许多城市的主要环境问题^[1]。目前,我国年废水排放总量约在400多亿t,每年排放干污泥约为550~600万t,且不断增加^[2]。到2010年,中国城市化率将达40%,城镇人口总量将增至6.7亿人。按每人每天产生50g(干物质)污泥计算,城市污泥排放量每年将达1200万t(干物质)以上^[3]。因此,如何合理、经济有效地处置数量巨大的污泥已成为非常紧迫的任务。

1 污泥处理与处置方法

目前,国内外污泥处理与处置方法主要为土地利用、填埋、焚烧和投海。因填埋填坑中有害物质通过雨水浸蚀和渗漏污染地下水,而焚烧成本高,易产生大气污染等^[2]。因此,填埋、焚烧和投海近年日渐减少或被禁止。城市污泥一般含有较高的有机质和矿质营养(有机质300~600 g·kg⁻¹、氮10~40 g·kg⁻¹和磷6~15 g·kg⁻¹)^[4]。据计,我国29个城市污泥(不包括工业污泥)有机质平均含量为384 g·kg⁻¹,全氮、全磷和全钾分别为27、14.3 g·kg⁻¹和7 g·kg⁻¹;有机质、全氮、全磷比纯猪粪分别高出1/3~2/3(全钾比纯猪粪低1/3),有机质含量呈逐年增加趋势^[5]。因此,城市污泥直接或经过干燥、堆肥等方式形成生物污泥(biosolids)应用于不同土地类型已是积极有效的处置方式,尤其一般林场、森林等地区污泥环境容量较大,在集约林地、城市园林、绿地及开垦荒地可避开食物链,对人类健康无害,施用安全^[2,4,6,7]。

污水中有50%~80%以上的重金属浓缩于剩余污泥中,其含量约为干污泥的0.5%~2%,某些情况下重量达4%。因而,污泥土地利用前须进行无害化预处理,一般有好氧厌氧消化、加石灰稳定法、堆肥等处理方法,一则降低污泥体积,便于运输和施用;二则降低乃至去除病原菌、有毒有害有机物、重金属等^[2,8,9]。近年亦出现了辐照污泥的方法^[10]。而堆肥是目前许多国家处理污泥的主要方式^[4],其指在一定条件下(如pH、通气、水分、温度等),将污泥等有机废弃物进行好氧堆沤,使之转化成类腐殖质的过程。堆肥后,污泥物理性状改善,质地疏松,营养成分增加,病原菌、寄生虫卵等基本杀死,重金属有效态含量降低,有毒有机物降解率平均在60%以上。将污泥和垃圾或木屑、树皮、谷物壳、秸秆等复合堆肥,或与少量化肥制成污泥有机-无机复混肥,是近年开辟的新处理途径^[11~14]。研究表明,污泥堆肥较厌氧消化或未经处理的污泥,引起土壤微生物区系平衡变化最小,并对土壤保持长期效用^[11,15];污泥堆肥较厌氧消化、颗粒污泥而言,降低Fe、Mn、Ni和Zn有效性,但增加Cu有效性^[16];污泥与其他生物废弃物堆肥,可能降低地下水污染的风险^[17]。从安全性看,污泥堆肥用作园林肥料较农用更具广阔前景^[18]。总之,无害化的污泥进行土地利用,益于:(1)降低潜在的污染源;(2)利于提高土地生产力,降低对合成肥料的需求^[16]。

2 污泥在农田生态系统的应用

2.1 不同污泥类型和用量对作物生长及产量的影响

施用城市污泥(有机质含量36.6%)可增加棉花产量和叶片K和P含量^[19],但工业污泥有降低棉花生物及经济产量的趋势^[20]。活性污泥应用后,燕麦生长和营养吸收均增加^[21]。污泥复合肥(N-P₂O₅-K₂O=8-8-6)与等养分化肥相比,小麦最大增产达17%^[22];污泥与化肥复混肥作底肥,冬小麦株高增加2~4 cm,千粒重提高0.1~1.7 g,增产7.6%~20.2%^[23];污泥和城市垃圾混合堆肥用量37.5 t·hm⁻²和75 t·hm⁻²,春小麦产量较对照分别增加54.1%和14.7%^[24]。辐照污泥(5 kGy剂量的γ射线,氮43 g·kg⁻¹)可增加小麦和水稻产量11%~27%^[10]。制革脱毒污泥(有机质50.1%)促进水稻生长,水稻苗期叶绿素含量、生物量(鲜重)和产量分别增加了17.3%、17.96%和16.95%,但原始制革污泥则抑制水稻生长,产量较对照下降27%^[25]。另外,施用污泥和垃圾复混肥料(N+P₂O₅+K₂O=30%)后,油菜产量、叶绿素含量均明显提高^[26];堆肥污泥对油菜籽出苗有少许影响,但低量污泥(<150 t·hm⁻²)可促进油菜幼芽发育;而且对大麦和中国甘蓝的产量为正效应^[27]。在碱性退化土壤上,累积施用污泥(全N 22.2 g·kg⁻¹)可使大麦从发育初期乃至抽穗期的干物质生产和叶蛋白质含量提高,产量显著增加^[28]。堆肥污泥(有机质58.3%)加入泥炭,增加椰菜营养及重金

属含量,且泥炭与30%的堆肥污泥混合,使椰菜产量最高^[29]。热喷污泥复混肥(总养分26%)施于青菜,比等养分的无机复合肥多增产38%,青菜体内水溶性总糖等约提高30%^[30]。热处理污泥颗粒(含水分9.5%)施于灌溉玉米,供应的全N、P含量约为无机肥料(尿素、磷酸氢二铵和KCl混合肥料)的45%,污泥与无机肥料对玉米产量和质量等效^[31]。可见,污泥普遍促进作物生长,益于提高生产力,且同一作物或不同作物因污泥类型和施用方式的差异而增幅不一。

研究表明,污泥施用量至关重要,但污泥施用量的界限因不同作物品种及土壤条件而异。譬如,青菜产量随污泥用量增加而增加,但污泥用量高于40 t·hm⁻²时,青菜体内的硝酸盐含量超过国标,达80 t·hm⁻²后产量明显下降,Cd含量超标^[32]。污泥施用适量,水稻返青加快、分蘖增多,产量提高,但污泥施用过量超过6000 kg·hm⁻²,秧苗生长及返青受影响;达15000 kg·hm⁻²以上,后期分蘖难以控制,生育延迟,产量显著下降;达24000 kg·hm⁻²以上,秧苗返青率降至80%左右;用量增至48000 kg·hm⁻²,秧苗返青率不足50%^[33]。另有研究表明,对水稻而言适宜的污泥施用量应不高于45 t·hm⁻²^[34]。当厌氧和脱水污泥用量较大,N供应过量导致小麦倒伏及糖用甜菜和小麦品质变差^[35]。由此说明,污泥农用须根据作物品种及土壤条件的差异和作物对N的需要,避免过量施用污泥。

2.2 农田生态系统重金属含量特征

研究表明,堆肥污泥应用于大麦和中国甘蓝,在土壤表层(0~20 cm)、大麦籽粒和中国甘蓝叶片都有重金属(Cu、Zn)积累^[27]。污泥施用后,Zn含量在土壤、小麦及蚜虫中随污泥增加而显著增加,而且Zn在迁移中有生物放大作用,蚜虫Zn含量是土壤的4倍,重金属从土壤-植物-节肢动物乃至捕食性动物的迁移可能有风险^[36]。污泥应用多年后,土壤表层因Cd含量及溶解性增加而导致大麦秸秆中Cd含量增加^[37];硬质小麦叶片中Pb和Cu浓度较高,叶鞘中Cd、Cr、Cu和Pb浓度较高^[38];小麦籽粒中N、P、Zn和Cu含量、糖用甜菜根N和Cu含量以及玉米籽粒中Cu含量均显著增加^[35]。污泥性状受处理方式的影响,没有消化处理的污泥,Cd是潜在有毒元素中最有害的元素之一,在玉米植株与籽粒中明显较高^[39]。

3 污泥在草地生态系统的应用

目前,城市污泥主要应用于人工草地,不仅增加生物量,而且改善土壤理化特性。譬如,干污泥(有机质42.97%~50.59%)施于无芒雀麦,3a地上生物量增幅分别为163%~203%、41%~120%和35%~106%;0~20cm土壤有机质、全氮、全磷和全钾含量显著增加,土壤容重减小^[40]。污泥(有机质35.58%)应用使早熟禾和结缕草生物量增加,早熟禾吸收富集Pb,污泥在15、30、60 t·hm⁻²低施入量时,结缕草吸收富集Cd、Cu和Zn^[41];污泥堆肥45 t·hm⁻²(干重)以下时,结缕草土壤中有机质、速效N、总N和总P分别比对照增加16%、78%、61%和140%,结缕草生物量提高1倍^[42]。污泥加入不同比例粉煤灰钝化后,高麦草干物重显著增加。随着粉煤灰加入量增加,高麦草地上部Ca、Mg和B的浓度增加,而K、Fe、Mn和Zn的浓度下降,且粉煤灰钝化后没有重金属毒害现象^[43]。

对黑麦草而言,施用14~70 t·hm⁻²的污泥堆肥(全N 16.1 g·kg⁻¹),黑麦草的地上地下生物量和叶绿素含量明显提高;与不施化肥和施用化肥的对照相比,污泥堆肥可增加土壤有效态氮、磷含量^[44];而施用污泥与化肥的复合肥可促进其根系发达及叶片对N的吸收,并提高草密度、覆盖度以及质量^[45]。另外,付华等^[46]施用污泥于黑麦草也有类似结论,即随污泥施用量(0.5~8 kg·m⁻²)增加,黑麦草盖度、密度、地上、地下生物量及叶绿素含量均显著增加,其地上部组织N、P、K、Mg、Fe、Zn和Cu含量增加,Mn含量减少,Ca和重金属元素Hg、Ni、Pb、As含量与对照间差异不显著;污泥适宜施用量为2.0~4.0 kg·m⁻²。同样,紫羊茅、马尼拉草和白三叶等施用污泥堆肥后,均获得良好生长响应,土壤理化性质明显改善;长期施用没有镉、汞毒害现象^[47]。

研究表明,施用污泥(全N 6.5%)改善紫花苜蓿生长,并随污泥用量增加对紫花苜蓿茎中Fe、Cu和Zn含量没有影响^[48]。污泥表施可增加荒漠草原生物量产量^[49],并通过叶面积的增加而发挥有利作用,污泥春季应用光合速率较夏季高,但植物对污泥的响应却因灌溉制度和物种而异^[50]。

4 污泥在森林生态系统的应用

目前,森林生态系统的污泥应用远不及农田土地类型。研究表明,污泥施用可提高森林生产力并改善土

壤特性^[51]。污泥(仅镉不符合农用标准,全N 17.1 g·kg⁻¹)施用于6年生毛白杨人工纯林,平原林和丘陵林地的毛白杨胸径净生长量分别比对照提高5.3%~17.6%和1.2%~5.8%,并以污泥施用量75 kg·株⁻¹效果最佳^[52]。在杨树、泡桐、油松林木上施用干污泥,林木高度和直径都随污泥施用量增加而增加^[53]。污泥堆肥(有机质44.39%)施于榆树较对照树高增加11%~25%,地径粗增加19%~50%^[54];污泥(用量300和600 kg N hm⁻²)施用在6龄松树上,污泥显著促进树木生长,5a后被处理的树木最大枝条的直径较对照几乎超过了1 cm^[55];污泥施用(不加或混合石灰)均导致柳树茎生物量大幅增加^[56, 57],且石灰污泥处理的土壤表层(0~10cm)pH约高2个单位,这对低pH土壤及土壤酸化尤其有利^[57]。污泥和垃圾堆肥(全N 23.6 g·kg⁻¹),明显促进苗木生长,提高叶片叶绿素含量,延长叶片生长期^[58]。

施用制革污泥(有机质48.88%)的杉木生长轮较未施污泥的宽,心材颜色深,且早材管胞的长度、弦向直径、射线频率和高度、壁的厚度均较高,并使杉木生长最快时期提前3~4a,缩短主伐周期,并促进具有较强吸收Cr能力的凤尾蕨生长^[59]。不同类型污泥(液态和脱水)应用在50a的欧洲赤松(*Pinus sylvestris L.*),土壤养分增加,松针N、Ca及Mg含量提高,土壤pH及铵态氮和硝态氮浓度显著提高,并降低腐殖质层的C/N比,特别是对pH和C/N比的影响在11a后仍显著,且脱水污泥较液态污泥有显著的长期效应,生态效益更为有利^[60]。由此说明,污泥对树木应用的生态环境效应与污泥的类型和用量相关。

在丘陵地区,施用污泥(有机质31.9%)可有效促进树木生长发育,增加株高和地径,对林中的灌、草层植被也有促进作用^[61]。特别是典型丘陵地区缺乏营养的林地施用污泥后,对森林生态系统营养循环可能产生的环境影响,以及商业林区的污泥持续利用,都日益引起广泛关注和重视^[62]。

5 污泥对土壤生态系统的影响

5.1 污泥对土壤理化性质的影响

由于污泥有机氮高,C/N比低,使污泥矿化迅速及肥效较高,施入土壤后,有机物分解的同时也部分形成土壤腐殖物质^[63]。研究表明,施用污泥对土壤有机质、全N及土壤理化性状如土壤电导率和速效N、P、K含量提高有利^[20, 64, 65],并增加土壤有效N、土壤全C及pH^[66]。特别是厌氧消化污泥和活性污泥应用5周后,土壤无机N从25增至>60 kg·hm⁻²^[22]。污泥施用20a后(用量100 t·hm⁻²·a⁻¹,每两年施用),土壤有机碳约为对照的2.5倍;污泥施用近6a,有机碳含量和微生物生物量均较对照高^[67];好氧和厌氧消化的污泥施用10a,中高用量(800、1200 kg N hm⁻²·a⁻¹)导致土壤有机质、腐殖质及腐殖酸显著增加,但对碳水化合物、团聚体稳定性无影响^[68]。低用量污泥的长期效应对土壤有机质及其组分亦没有影响^[69]。

污泥堆肥施入土壤后使其结构系数、水稳定性减小,促进雨水入渗、减少水土流失^[54];并增加土壤持水能力、团聚稳定性和阳离子交换量,减少土壤容重^[1]。盐化土施用污泥后,土壤改良主要是由于碱化度减少,而使非盐化土壤结构改善的主要因子是有机碳和碳水化合物含量增加,即污泥应用于土壤通过增加团聚体稳定性和降低土壤容重来改善土壤结构^[70];石灰性土施污泥后,提高土壤生物化学特性,降低土壤pH,增加土壤盐分,并随其用量的增加,累积C矿化、溶解有机C、腐殖质和棕黄酸C及微生物生物量C均显著增加^[71]。如污泥和煤灰以低至中等用量混合施用,没有显著的金属吸收或淋溶,利于土壤修复^[72]。

污泥用于森林土壤可显著提高营养元素和相关重金属含量,如一次大量施入,可维持相当长时间的高浓度;在植被很少、质地极粗的土壤上一次施入厚达25 cm的污泥,15a后大量养分特别是钙和镁等移动性很大,可移至较深土层^[73]。其原因可能是森林土壤pH较低,显著增加金属微量元素的有效性^[74]。但在欧洲赤松成熟林施用污泥后,有机土壤层可溶性铵和磷酸盐及Bs层的磷酸盐明显增加,归因于污泥中养分淋溶,且污泥施用后土壤pH或有效C增加,促进土壤矿化过程,导致N、P淋失提高,从而可能影响营养循环以及森林土壤的营养损失^[62]。

在矿山恢复系统中,施用污泥可提供有机碳并改善土壤肥力,改良矿山废弃地的理化性质和防治水土流失,利于迅速有效地恢复植被,并提高矿山废弃地微生物的活性^[75, 76]。特别是污泥应用于铅/锌尾矿的垦植,不仅增加全碳含量及N、P、K含量,而且降低尾矿基质的Zn、Pb、Cd全量及Pb和Cd的DTPA(二乙烯三胺五

乙酸)提取量,继而降低了长喙田菁和田菁植物器官对 Zn、Pb、Cd 的吸收和积累^[77]。

5.2 污泥对土壤酶和微生物的影响

污泥施用于土壤生态系统,对土壤质量的影响具有双重性或结论不一,这种差异可能取决于污泥重金属含量和污泥分解速率。一方面,污泥施用使土壤微生物数量和活性增加,土壤中脲酶、过氧化氢酶、多酚氧化酶及中性磷酸酶的活性增加,从而改善土壤结构,譬如污泥与化肥配施苇状羊茅后,较单施化肥提高砾石风化物微生物总数量 4~23 倍,脲酶活性 1.8~2.8 倍,生物量碳 0.3~2.4 倍,说明污泥利于砾石山微生物区系的建立^[78~81];另一方面,无机或/和有机污染物富积对土壤微生物区系有负面影响,土壤遭受金属污染,譬如污泥中较高浓度 Zn,使微生物活性、生物量及生物量 C 降低^[82~86];污泥分解时期,污泥中一些重金属可能负面影响土壤酶活性(b-葡萄糖苷酶,碱性磷酸酯酶,硫酸酯酶和脲酶),污泥中高 N 含量亦可能导致酶抑制^[87]。

土壤微生物活性对生物化学过程和农用化学药品转化过程具有重要作用,并直接影响生态系统稳定和土壤肥力。因此,微生物参数被认为是土壤遭受胁迫和化学污染引起土壤环境改变的敏感指标^[15, 88],其中土壤呼吸、微生物生物量及固氮被认为是研究土壤微生物区系最常用的参数^[11, 83, 89],用于评价环境污染物对土壤肥力及农业生产力的影响^[82]。研究表明,污泥连续应用后,随污泥增加,土壤微生物生物量 C、N、基础呼吸、代谢商及酶活性均随之增加,并与污泥用量呈正相关^[90]。

不同城市污泥类型引起的响应不同,厌氧消化、堆肥及未经处理的原始污泥应用于典型灰色森林土并种植春大麦,堆肥污泥较无污泥的对照可增加土壤微生物生物量约 1.9~4.4 倍,土壤基础呼吸约 2.3~6.3 倍及固氮活性约 2.1~35 倍,而厌氧消化污泥对微生物量及其活性没有显著影响;原始污泥显著降低了固氮活性,而堆肥污泥对土壤微生物生物量及其活性有利^[11]。另外,污泥增加导致动物区系活性增加,这可从排泄物富积及土壤总孔隙度增加显示出来^[91]。反之,在污泥中加入蚯蚓,则有利于降低污泥中 Cu 和 Cd 的生物利用度,并显著改善土壤物理属性和有效养分^[92]。

施用污泥对土壤固氮菌影响较小,而对硝化菌影响较大^[93]。污泥施入农地,锌毒害降低细菌多样性^[94]。污泥应用在松类及混交再生林土壤,对链球菌指示细菌数量没有显著影响,而总大肠菌群体随时间和污泥处理水平而变,因此污泥应用可能有一些细菌污染的风险,并取决于污泥类型和森林环境^[95]。污泥对根瘤菌种群的大小和脱氢酶活性的影响明显有利,且与土壤 pH、污泥类型及用量和现存金属含量有关^[96]。另外,在旱地和水田施用污泥复混肥与等量化肥相比,明显促进硝化细菌和好气性纤维素分解细菌的繁衍,对好气细菌、真菌、放线菌和氨化细菌也略有促进作用,并随污泥复混肥中污泥含量的提高而增强,但对施用多少量其激活作用最大或使用多少量和多长时间才会影响土壤微生物的生化活性或导致土壤和农产品的重金属污染,尚待深入研究^[14]。

6 污泥应用的环境效应

一般情形下,污泥土地利用产生的主要环境问题可归结为:重金属污染;病原物污染;氮磷过剩引起的地表水和地下水污染等^[63]。因污泥用量、时期以及环境影响因子如土壤特性等不同,污泥应用产生的环境效应研究结论亦不一。

研究表明,污泥可能会因导致营养失衡及有毒元素积累和淋溶而给环境造成危险^[97, 98]。污泥携带重金属滞留土壤,吸附土壤含水氧化物、粘土及有机质,并以不溶性盐形式或污泥残留微粒存在^[99]。在灌区沙土壤施用污泥,土壤表层 20cm 遭受重金属污染,Zn 和 Cu 积累,因受制于有机质而仅限于表层土壤,对地下水未造成污染^[100]。污泥堆肥利用使土壤 Cd 含量比对照增加了 12.95%~154.48%,而土壤 Pb、Cu、Zn 含量变化甚微^[54]。通过模拟 20a 的污泥施用,表明 Cd 对地下水污染风险微小,但土壤表层高 Cd 含量将导致作物吸收 Cd 的风险较大^[101]。污泥修复的土壤,Cu、Zn 淋溶数量可观^[102];土壤 Cu 和 Zn 有效态增加,Cu 和 Zn 全量和有效量在蚯蚓粪便中均高于对照土壤^[103],而且因污泥修复增加了土壤 P 有效性,可能导致水富营养化的危害^[35]。

污泥长期施用可增加土壤重金属总量^[41]。污泥处理 40a 后的土壤剖面,在表土和底土中 Zn、Cu、Cd 和

Pb 的平均含量分别为 $79.3, 32, 0.29 \text{ mg} \cdot \text{kg}^{-1}$ 和 $1.15 \text{ mg} \cdot \text{kg}^{-1}$, 在污泥土壤剖面 AB 层, 所有金属含量均高于对照, Pb 和 Cu 具有相对高的迁移性, 这些足以引起健康和植物性毒素的风险, 因为 Cd 可能进入食物链, 而 Zn、Cu 和 Pb 可能减少土壤生产力或污染浅蓄水层^[104]。土壤特性如土壤有机质和土壤 pH 可能是影响金属迁移的主要因素。譬如可溶性有机质增加使 Cu 迁移增加, 特别是高 pH 的沙土^[105]。

然而, 在酸性土壤施用污泥, 重金属的迁移可以忽略^[106]。污泥中高有机质含量及相应低的 pH, 可能利于金属-有机质的络合, 从而减少重金属迁移以及危害植物的毒性^[104]。奶牛场污泥应用于草地, 短期内(4a)不会导致重金属有害积累^[107]。污泥施用于粘壤土, Cd 的长期迁移(41a)表明, 表层土壤中有 92% 的污泥 Cd, 而 7% 滞留在底土上层 17cm, Cd 淋失没有造成高风险^[37]。而在杨树、泡桐、油松等林木上施用污泥亦未引起地下水污染, 重金属均累积于土壤表层, 未向下层转移^[53]; 在丘陵区树木土壤施用污泥后, 除 Pb 含量随污泥用量增加而增加外, 其它重金属增加不显著^[58]。另外, 园林绿地施用污泥堆肥后, 表层土(0 ~ 20 cm) Cd、Pb、Cu、Zn、Ni、Cr 的含量均明显提高, 但径流水和渗滤水中 6 种重金属元素含量都很低^[108]。污泥用于森林土壤 15a 后, 重金属一般未移动至 20cm 以下, 由于施用污泥会引起酸化作用增强, 使 Cd 和 Pb 的溶解度增加而移动至 17 ~ 27 cm, 但重金属仍停留在土壤表层, 基本未向下层移动^[73]。大豆施用污泥后, 在土壤深 2m 处取样, $\text{NO}_3\text{-N}$ 浓度略高于对照; 地下水 $\text{NO}_3\text{-N}$ 浓度有所增高, 但差异很少显著^[109]。不过, 污泥施用增加 CO_2 、 N_2O 和 CH_4 向大气排放通量, 污泥最高用量(是玉米连作 4 次 N 含量的 8 倍)较对照分别增加 220%、320% 和 165%, 但有一部分 C 被固定在土壤中^[110]。

7 研究展望

目前, 污泥作为一种有机肥料已成为普遍措施。尽管污泥利用会给植物-土壤系统带来重金属污染风险, 但在这些元素浓度低的待修复土壤中, 可能给植物带来有益的微量元素。随着城市化进程加快和人口剧增, 我国污泥将大量增加, 土地利用将是我国污泥处置的主要方法。尤其我国 K 肥严重缺乏, P 肥品位较低且大多难以开采, 加之大面积中低产田需改良, 废弃矿山需复垦。因此, 将经过稳定化、无害化处理后的污泥进行土地循环利用, 应该是我国污泥处置利用、改良土壤以及防治水土流失的一种具有前景的发展途径。不过, 与国外相比, 我国污泥农林利用率较低, 对城市污泥资源化的理论研究与实践均相差甚远。故而, 今后污泥处理及利用的研究展望具体如下:

(1) 加强污泥土地利用后植物-土壤之间生态环境效应基础理论的系统研究。譬如, 污泥应用后提高植物生产力的生理生态机制; 污泥应用后土壤环境中重金属行为以及植物重金属的迁移机理; 污泥重金属可能引起的生境污染; 环境改变(譬如土壤质量)导致的植物多样性和生产力的变化(譬如杂草侵入)等。

(2) 加强污泥施用对生态环境长期影响的定位实验研究和长期定位监测。譬如, 污泥中金属微量元素在深土层及地下水淋溶的长期研究; 污泥有机质对土壤化学、生物和物理特性的长期影响; 污泥矿化对营养循环的长期影响。特别是土壤环境条件(如温度、pH 和土壤水分含量等)变化时, 土壤微生物区系受之影响, 而土壤微生物是营养循环(C、N 及 P)中的重要部分, 对土壤有机质降解起主要作用, 因此污泥中综合有毒物对土壤微生物区系的影响已成为当前研究领域的一个热点。

(3) 由于污水来源不同和处理技术各异, 其成分、含量及特性差别很大, 因此有必要研究不同类型及不同用量的污泥有机肥料或复合肥的生态安全适宜性, 并有必要区分不同类型污泥营养有效性的差异, 以便避免不充足或者过量的应用比例。正确评价污泥营养的有效性, 有可能会降低对商业肥料供应的需求。

(4) 增强城市污泥土地利用潜在的风险评价研究, 尤其是城市污泥中有机污染物质的累计风险以及病原体的危害研究。

(5) 寻求经济有效的污泥处理利用的新技术或集成技术示范研究。目前污泥稳定化处理技术主要有热喷、辐射、堆肥、碳化污泥处理等技术, 今后须改进或创新污泥处理工艺或技术, 减少污泥中有害物质的含量, 提高污泥质量, 并加强污泥处理技术—污泥应用量化技术—产业化的集成技术示范研究。

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